



A novel ion-exclusion chromatography–mass spectrometry method to measure concentrations and cycling rates of carbohydrates and amino sugars in freshwaters

Hornák, Karel ; Pernthaler, Jakob

Abstract: The concentrations of free neutral carbohydrates and amino sugars were determined in freshwater samples of distinct matrix complexity, including meso-, eu- and dystrophic lakes and ponds, using high-performance ion-exclusion chromatography (HPIEC) coupled to mass spectrometry (MS). In contrast to other methods, our approach allowed the quantification of free neutral carbohydrates and amino sugars at low nM concentrations without derivatization, de-salting or pre-concentration. New sample preparation procedures were applied prior to injection employing syringe and hollow fiber filtration. Analytes were separated on a strong cation exchange resin under 100% aqueous conditions using 0.1% formic acid as a mobile phase. To minimize background noise in MS, analytes were detected in a multiple reaction monitoring scan mode with double ion filtering. Detection limits of carbohydrates and amino sugars ranged between 0.2 and 2nM at a signal-to-noise ratio >5. Error ranged between 1 and 12% at 0.5-500nM levels. Using a stable isotope dilution approach, both the utilization and recycling of glucose in Lake Zurich was observed. In contrast, N-acetyl-glucosamine was equally rapidly consumed but there was no visible de novo production. The simple and rapid sample preparation makes our protocol suitable for routine analyses of organic compounds in freshwater samples. Application of stable isotope tracers along with accurate measures of carbohydrate and amino sugar concentrations enables novel insights into the compound in situ dynamics.

DOI: <https://doi.org/10.1016/j.chroma.2014.09.007>

Posted at the Zurich Open Repository and Archive, University of Zurich

ZORA URL: <https://doi.org/10.5167/uzh-106911>

Journal Article

Accepted Version

Originally published at:

Hornák, Karel; Pernthaler, Jakob (2014). A novel ion-exclusion chromatography–mass spectrometry method to measure concentrations and cycling rates of carbohydrates and amino sugars in freshwaters. *Journal of Chromatography. A*, 1365:115-123.

DOI: <https://doi.org/10.1016/j.chroma.2014.09.007>

Accepted Manuscript

Title: A novel ion-exclusion chromatography-mass spectrometry method to measure concentrations and cycling rates of carbohydrates and amino sugars in freshwaters

Author: Karel Horňák Jakob Pernthaler



PII: S0021-9673(14)01387-9
DOI: <http://dx.doi.org/doi:10.1016/j.chroma.2014.09.007>
Reference: CHROMA 355796

To appear in: *Journal of Chromatography A*

Received date: 21-5-2014
Revised date: 11-8-2014
Accepted date: 2-9-2014

Please cite this article as: K. Horňák, J. Pernthaler, A novel ion-exclusion chromatography-mass spectrometry method to measure concentrations and cycling rates of carbohydrates and amino sugars in freshwaters, *Journal of Chromatography A* (2014), <http://dx.doi.org/10.1016/j.chroma.2014.09.007>

This is a PDF file of an unedited manuscript that has been accepted for publication. As a service to our customers we are providing this early version of the manuscript. The manuscript will undergo copyediting, typesetting, and review of the resulting proof before it is published in its final form. Please note that during the production process errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.

1 **A novel ion-exclusion chromatography-mass spectrometry method to measure**
2 **concentrations and cycling rates of carbohydrates and amino sugars in freshwaters**

3

4 Karel Horňák^{*}, Jakob Pernthaler

5

6 *Limnological Station, Institute of Plant Biology, University of Zurich, Seestrasse 187, CH-8802*

7 *Kilchberg, Switzerland*

8

9 ^{*}Corresponding author. Tel.: +41 44 634 9236; fax: +41 44 634 9225

10 *E-mail addresses:* khornak1@gmail.com (K. Horňák), pernthaler@limnol.uzh.ch (J.

11 Pernthaler).

12

13 *Keywords:*

14 Carbohydrates; Amino sugars; Ion-exclusion chromatography, Mass spectrometry, Stable

15 isotope tracers; Freshwater

16 **Abstract**

17 The concentrations of free neutral carbohydrates and amino sugars were determined in
18 freshwater samples of distinct matrix complexity, including meso-, eu- and dystrophic lakes
19 and ponds, using high-performance ion-exclusion chromatography (HPIEC) coupled to mass
20 spectrometry (MS). In contrast to other methods, our approach allowed the quantification of
21 free neutral carbohydrates and amino sugars at low nM concentrations without
22 derivatization, de-salting or pre-concentration. New sample preparation procedures were
23 applied prior to injection employing glass fiber and hollow fiber filtration. Analytes were
24 separated on a strong cation exchange resin under 100% aqueous conditions using 0.1%
25 formic acid as a mobile phase. To minimize background noise in MS, analytes were detected
26 in a multiple reaction monitoring scan mode with double ion filtering. Detection limits of
27 carbohydrates and amino sugars ranged between 0.2 – 2 nM at a signal-to-noise ratio >5.
28 Error ranged between 1 - 12 % at 0.5 - 500 nM levels. Using a stable isotope dilution
29 approach, both, the utilization and recycling of glucose in Lake Zurich was observed. In
30 contrast, N-acetyl-glucosamine was equally rapidly consumed but there was no visible *de*
31 *novo* production. The simple and rapid sample preparation makes our protocol suitable for
32 routine analyses of organic compounds in freshwater samples. Application of stable isotope
33 tracers along with accurate measures of carbohydrate and amino sugar concentrations
34 enables novel insights into the compound *in situ* dynamics.

35 1. Introduction

36 Dissolved organic matter (DOM) in freshwater ecosystems consists of a wide array of
37 compounds that are in parts linked to processes mediated by aquatic microbes [e.g., 1].
38 DOM is a key nutrient and energy source for consumers, particularly for heterotrophic
39 bacteria [e.g., 2,3]. Major sources of DOM in freshwaters are autochthonous primary
40 production and the allochthonous input from the terrestrial environment [4-6].

41 Carbohydrates and amino sugars often constitute a prominent fraction of the labile
42 DOM fraction in aquatic systems [e.g., 7,8]. Yet the quantification of dissolved carbohydrates
43 and amino sugars is analytically challenging due to their typically low *in situ* concentrations
44 at nM levels [9], the chirality at many carbon atoms and the existence of anomers resulting
45 in a large number of isomers. Diverse colorimetric and chromatographic approaches have
46 been used for the determination of environmental carbohydrates. However, the application
47 of colorimetric methods [10-12] has been limited by sensitivity, poor detection limits ($\sim\mu\text{M}$)
48 and by time-consuming derivatization steps. Similarly, chromatographic methods employing
49 borate complexes [13,14], reversed-phase chromatography with DNS and p-AMBA [15]
50 derivatives or gas chromatography with ester or acetyl derivatives [16,17] all require difficult
51 sample preparation and their detection limits range between 100 nM to 20 μM .

52 To date, only high-performance anion-exchange chromatography with pulsed
53 amperometric detection [HPAEC-PAD, 18] allows the detection of carbohydrates and amino
54 sugars with sufficient sensitivity (2-10 nM) and without laborious derivatization procedures.
55 The HPAEC-PAD method has been employed for the quantification of neutral sugars [18],
56 amino sugars [19], various disaccharides [20] or a combination of those compounds [21] in
57 samples from various aquatic habitats. Despite being widely applied, the sensitivity and
58 specificity of the HPAEC-PAD method may be compromised by measurement interferences

59 with salt anions and a complex sample matrix [e.g., 19,22]. To remove possible contaminants
60 (e.g. amino acids) -resulting in unspecific signals- clean-up columns have been commonly
61 employed prior to HPAEC [e.g., 21,22]. Thus, alternative detection techniques might help to
62 improve the specificity of the analysis.

63 Liquid chromatography (LC) coupled to mass spectrometry (MS) is a powerful
64 approach for the determination of carbohydrates in freshwater samples [e.g., 6,23].
65 Methods employing mass spectrometric detection are particularly suitable due to their
66 potential for high sensitivity and specificity of the analyses. In contrast to detection by PAD,
67 molecular identity of the analytes can be confirmed by MS. In addition, different isotopes of
68 a given element can be distinguished, allowing for stable isotope pulse labeling studies to
69 track the distribution and cycling of specific organic compounds in aquatic systems.

70 The determination of microbial uptake rates of dissolved carbohydrates has been
71 predominantly based on radiolabeled assays [24-27] or cycling parameters of individual
72 compounds have been derived from differences in substrate concentration over time [e.g.,
73 28,29]. Occasionally, stable isotope tracers have been applied to examine the compound
74 dynamics *in situ* [e.g., 30,31,32]. A major advantage of the latter approach is that both, the
75 natural and the added stable isotope-labeled compounds can be monitored simultaneously
76 by MS. Furthermore, stable isotope dilution models [33,34] allow a simultaneous estimation
77 of compound utilization and production rates, which is important for understanding its
78 biogeochemical cycling.

79 The aim of this study was to develop and validate a rapid and sensitive LC-MS
80 protocol for the determination of dissolved free carbohydrates and amino sugars in diverse
81 freshwater habitats at their low nM concentrations along with a simple isotope dilution
82 approach to quantify the utilization and production rates of selected analytes *in situ*. Here,

83 we describe the novel method in detail and compare its performance with other approaches
84 for the analysis of the target compounds. Furthermore, we show that our approach allows to
85 simultaneously combine the accurate quantification of carbohydrate and amino sugar
86 concentrations with their cycling rates, thereby providing insight into their *in situ* dynamics.

87

88 **2. Methods**

89 *2.1. Reagents*

90 D-glucose (Glc), D-mannose (Man), D-galactose (Gal), D-fructose (Fru), L-fucose (Fuc),
91 L-rhamnose (Rha), L-arabinose (Ara), D-lyxose (Lyx), D-ribose (Rib), D-xylose (Xyl), cellobiose
92 (Cel), sucrose (Suc), maltose (Mal), N-acetyl-D-glucosamine (GlcNAc), N-acetyl-D-
93 galactosamine (GalNAc), N-acetyl-D-mannosamine (ManNAc), N-acetylmuramic acid
94 (MurNAc), N,N'-diacetylglucosamine (GlcNAc)₂, sucralose (Scl) were purchased from Sigma-
95 Aldrich at >98% purity grade. LC-MS grade formic acid (50%) and ammonia (25%) were
96 obtained from Sigma-Aldrich. LC-MS grade water (Carl Roth) was used for the preparation of
97 solvents and aqueous reagents. D-[UL-¹³C₆]-glucose (Cambridge Isotope Laboratories) and N-
98 acetyl-D-[UL-¹³C₆]-glucosamine (Omicron Biochemicals) were used to determine the
99 production and utilization rates of Glc and GlcNAc in Lake Zurich.

100 *2.2. Sample collection and preparation*

101 Water samples of 500 ml were collected from the large mesotrophic prealpine Lake
102 Zurich (406 m above sea level, area 65.06 km², max. depth 136 m, mean depth 51.7 m,
103 residence time 440 d, 47°17'N, 8°36'E, Switzerland, [35]) into acid-washed glass bottles and
104 were kept at *in situ* temperature in the dark. Additionally, a small eutrophic lake, Hüttensee
105 (658 m above sea level, area 0.165 km², max. depth 13.3 m, mean depth 6.3 m, residence
106 time 120 d, 47°14'N, 8°38'E, water quality parameters obtained from the Office of Waste,

107 Water, Energy and Air, Canton Zurich, <http://www.awel.zh.ch/internet/audirektion>
108 /awel/de/wasserwirtschaft/messdaten/see_qualitaet.html) and a small dystrophic peat bog
109 located in the Unterrifferswilermoos area (pH = 4.8, water color – 10.93 mg tannic acid
110 equivalents l⁻¹, 47°14'N, 8°30' E) were sampled once to evaluate the performance of the
111 method on water with a different matrix complexity. Samples were processed by the newly
112 developed LC-MS protocol within 1 h after the sampling. Subsamples of 5 ml were collected
113 using a gas-tight glass syringe (Hamilton) and subsequently filtered through a glass fiber
114 syringe filter (pore size 0.2 µm, diameter 25 mm, Tisch Scientific). Prior to the filtration, the
115 syringe and the filters were first washed once with 2M HCl, then 3 times with sterile
116 deionized water and finally with LC-MS grade water. Filtrates of 1.5 ml were collected into
117 HPLC vials (volume 2 ml, Glastechnik Graefenroda). First 1 ml of the filtrate was discarded.
118 Samples were processed immediately upon filtration or stored at -20°C until further
119 analyses. Visual water color was measured spectrophotometrically at 440 nm and expressed
120 as tannic acid equivalents [36].

121 2.3. Ion-exclusion chromatography

122 Carbohydrate analyses were performed with a HPLC system (1260 Infinity series,
123 Agilent Technologies) equipped with a degasser, a binary pump, an autosampler and a
124 column oven. Analytes were separated on a Supelcogel C-610H column (300 x 7.8 mm i.d., 9
125 µm particle size, sulfonated polystyrene/divinyl benzene, counter ion H⁺, Supelco) protected
126 by a column pre-filter (pore size 0.45 µm, Brechbühler). The mobile phase flow rate was 700
127 µl min⁻¹; it consisted of a gradient of 0.1% formic acid (solvent A) and water (solvent B). The
128 following elution gradient was applied: A from 20 to 50% in 6.9 min, then 50 % A was
129 maintained for 7.4 min, then reduced to 20% A in 0.1 min and this was maintained for 7.6
130 min. The column temperature was constant at 60°C. Samples were kept at 4°C during

131 analyses and the injection volume was 50 μ l. An internal standard (50 nM sucralose,
132 injection volume 50 μ l) was injected from an external vial 4 min after the sample injection by
133 a custom-defined injection sequence. Prior to the first sample injection, the column was
134 equilibrated for 45 min at 20% A until a stable back pressure level was reached, followed by
135 2 blank injections using the gradient elution as detailed above. To ensure the quality of the
136 separation, retention times of the analytes and the peak area of the internal standard were
137 monitored during analyses.

138 *2.4. Mass spectrometry*

139 Analytes were detected on an API 5000 triple quadrupole mass spectrometer (AB
140 Sciex) equipped with an electrospray ionization probe. Using a multiple reaction monitoring
141 (MRM) scan mode, carbohydrates and amino sugars were detected in the negative and
142 positive ion mode, respectively. Depending on the analyte retention time, only relevant
143 transitions (pairs of precursor and product ions) were monitored at the same time. If
144 required, different MS conditions were applied within a single run to increase the sensitivity
145 of the analysis. The software Analyst (version 1.6.1, AB Sciex) was used for the data
146 acquisition. Measurements of the extracted ion chromatographs (XIC) for each of the
147 monitored transitions were further quantitatively analyzed using the MultiQuant software
148 (version 2.1, AB Sciex). Peak areas (in arbitrary units) were determined by integration
149 algorithms implemented in MultiQuant. The signal to noise ratio threshold was 5.

150 *2.5. Compound optimization assays*

151 Aqueous stock solutions (1 mM) of each compound were stored at -20°C. For the
152 compound optimization assays, 50 to 500 nM working solutions were prepared from the
153 stock solutions. Tandem mass spectrometry (MS/MS) analyses were applied to obtain
154 compound-specific transitions with their specific declustering potential (DP), entering

155 potential (EP), collision energy (CE) and cell exit potential (CXP) parameters (Table 2). For all
156 transitions EP was set to 10 eV. Firstly, compound working solutions were directly injected
157 into the MS by a syringe pump (Harvard) at a flow rate of $10 \mu\text{l min}^{-1}$. Mass spectra (Q1 scan
158 mode) were recorded to confirm the presence of either the deprotonated molecule $[\text{M-H}]^-$ in
159 the negative ion mode (neutral carbohydrates) or the protonated molecule $[\text{M+H}]^+$ in the
160 positive ion mode (amino sugars). Additionally, product ion mass spectra (product ion scan
161 mode) were acquired from the precursor ion ($[\text{M-H}]^-$ or $[\text{M+H}]^+$). For all scans, the lower m/z
162 threshold was set to 50. To optimize ion-spray voltage (IS), temperature (TEM) as well as
163 curtain (CUR), collision (CAD) and ion source gas (GS1 and GS2) parameters, $10 \mu\text{l}$ of a
164 compound working solution was injected in 0.05% formic acid as a mobile phase at a flow
165 rate of $700 \mu\text{l min}^{-1}$. Signal intensities of the transitions were monitored in the multiple
166 reaction monitoring (MRM) scan mode and compared over a defined range of ion source
167 parameters. Nitrogen was used as curtain, collision and ion source gas. After optimization
168 the ion source parameters for carbohydrates were set as follows: IS = -2500 V, TEM = 600°C ,
169 CUR = 35, CAD = 7, GS1 = 57, GS2 = 58. For sucralose (internal standard): IS = -4000 V, TEM =
170 650°C , CUR = 37, CAD = 8, GS1 = 60, GS2 = 57 and for amino sugars: IS = 4000 V, TEM =
171 600°C , CUR = 45, CAD = 10, GS1 = 60, GS2 = 55. For each compound, the 2-4 transitions with
172 highest signal intensities were selected. For all transitions the dwell time was set to 100 ms.
173 Transitions will be denoted as in the following example: a molecule with a precursor ion at
174 m/z 179 and a product ion at m/z 89 will be given as 179/89. Details on the transitions are
175 given in Table 2.

176 *2.6. Internal standard and response factor*

177 For quantitative studies, dilution series of aqueous stock solutions (0.5 – 500 nM) of
178 each analyte and 50 nM solution of sucralose (internal standard) were prepared and

179 analyzed as described in sections 2.3. and 2.4. Peak areas of both analyte and internal
180 standard were acquired by the MultiQuant software. For each compound, a response factor
181 was calculated as the ratio of the quotient of the analyte and the internal standard peak
182 areas to the quotient of the analyte and the internal standard concentrations. This response
183 factor was then applied to determine the concentration of the target compound in
184 experimental samples.

185 2.7. Incubation experiments with ^{13}C -labeled Glc and GlcNAc

186 Water samples from specific depths within the upper 40 m of the water column of
187 Lake Zurich were collected into acid-washed glass bottles (volume, 1 l) in 2013 during the
188 summer stratification and autumn overturn periods. First, the *in situ* concentrations of Glc
189 were assessed in order to yield comparable atom% excess ^{13}C values. Subsequently,
190 triplicate sample bottles were supplemented with D-[UL- $^{13}\text{C}_6$]-glucose (final concentrations,
191 20-30 nM) and N-acetyl-D-[UL- $^{13}\text{C}_6$]-glucosamine (final concentrations, 10-20 nM) and
192 compared to duplicated control bottles without substrate amendments. Samples were
193 incubated at *in situ* temperature in the dark for 4 h. Subsamples of 5 ml were taken at
194 intervals of 1 h. Concentrations of natural and ^{13}C -labeled Glc and GlcNAc were determined
195 as described in sections 2.3. and 2.4. Identical pairs of transitions obtained from the natural
196 and ^{13}C -labeled compounds (differences in m/z correspond to the number of ^{13}C atoms)
197 were used to determine the atom% excess ^{13}C [30]. The corresponding compound utilization
198 and production rates were calculated according to an isotope dilution model [33,34].

199

200 3. Results

201 3.1. Filtration performance and compound recoveries

202 Since carbohydrates and amino sugars are readily utilized by microbes and the typical
203 turnover times of these compounds are in the range of hours [e.g., 27] it is essential to stop
204 the biotic reactions by filtration to avoid sample degradation prior to the LC-MS analysis.
205 However, an uncontrolled release of extractable impurities may occur from such filtration
206 devices and interfere with the analysis of targeted analytes [e.g., 37], or analytes can be
207 selectively retained by the filter membranes [e.g., 38]. Therefore, the performance of
208 different membrane types was evaluated: First, ion signal intensities of selected
209 carbohydrate and amino sugar transitions obtained prior and after the filtration of LC-MS
210 grade water were evaluated by the new LC-MS method. Subsequently, the recoveries of
211 selected carbohydrates and amino sugars (0.2-10 nM solutions in LC-MS grade water) were
212 investigated by comparing concentrations before and after the filtration.

213 Large interferences were observed after filtering the LC-MS grade water through
214 filters of cellulose acetate (CA), nylon, polytetrafluoroethylene (PTFE), and polyvinylidene
215 difluoride (PVDF) membrane types (Table 1). Samples filtered through the glass fiber (GF)
216 and polyethersulfone (PES) filters revealed virtually no or only very minor interferences as
217 compared to the non-filtered LC-MS grade water. Virtually all bacteria and other microbes
218 and particles present in natural samples were removed by 0.22 μm PES filters, whereas only
219 96-98 % of total bacterial counts could be removed by 0.22 μm GF filters. This was
220 determined by flow cytometry and confirmed by fluorescence microscopy (data not shown).
221 The presence of bacteria in the GF filtrate, however, did not result in significantly different
222 concentrations of the analytes compared to samples filtered through PES filters (data not
223 shown). Nevertheless, PES syringe filters allow for more standardized sample filtration and
224 are, therefore, recommended for sample prefiltration. No significant differences in
225 compound concentrations occurred prior and after the sample filtration through PES

226 membrane. Recoveries of the analytes ranged from 95 to 104 % at 0.2-10 nM levels (Suppl.
227 Table 1). In addition, the hollow fiber PES filters allow for a very efficient and gentle
228 preparation of large quantities of particle-free water by minimizing the danger of artificial
229 enrichment with solutes due to the breaking of algal cells [39].

230 *3.2. Efficiency of separation and detection*

231 Our chromatographic setup allowed an efficient separation of a number of neutral
232 carbohydrates and amino sugars. The separation factor (α) was determined by comparing
233 the retention factors of two given peaks. Nearly a baseline separation of Glc from other
234 hexoses ($\alpha = 1.14$) was achieved (Fig. 1A). In contrast, Man, Gal and Fru coeluted ($\alpha = 1$), and
235 even modified chromatographic conditions did not improve their separation. Fuc and Rha
236 could be well resolved (Fig. 1A, $\alpha = 1.16$) as well as Xyl, Lyx and Rib (Fig. 1B, $\alpha = 1.08$ and
237 1.25, respectively). On the other hand, Ara partly coeluted with Rib ($\alpha = 1.03$) and Lyx ($\alpha =$
238 1.1). Of the tested disaccharides, Cel, Mal and Suc could be resolved although a complete
239 separation was not achieved (Fig. 1B, $\alpha = 1.12$ and 1.71, respectively). The amino sugars
240 GlcNAc, ManNAc and GalNAc were efficiently separated (Fig. 1C, $\alpha = 1.11$ and 1.38,
241 respectively). Even though MurNAc and (GlcNAc)₂ revealed almost identical retention times
242 ($\alpha = 1.06$), these compounds could be unambiguously distinguished by their specific
243 transitions (Fig. 1C). Similarly, analytes from different carbohydrate classes (i.e.,
244 disaccharides, hexoses, deoxy-hexoses, pentoses) could be accurately determined by MS
245 although the chromatographic conditions applied did not allow for their complete
246 separation. Due to the isomeric nature of the analytes (e.g. hexoses), identical transitions
247 were used for their detection. Unfortunately, the poor chromatographic resolution of Man,
248 Gal and Fru did not allow for their accurate quantification and the compounds could be only
249 partly differentiated based on the relative signal intensities of the transitions.

250 3.3. Method precision and detection limits

251 A linear detector response was observed for the studied analytes over a broad range
252 of concentration (0.2-500 nM). Limits of detection (LOD) was determined as the lowest
253 concentration at which the studied analytes yielded a signal to noise ratio of >5 [40].
254 Depending on the sample types LOD values varied between 0.5-2 nM and between 0.2-1 nM
255 for carbohydrates and amino sugars, respectively. Method precision, assessed as the relative
256 standard deviation (RSD), was calculated from replicate measurements (n = 6) of natural lake
257 water spiked with known concentrations of selected compounds. Limits of quantification
258 (LOQ) were defined as the lowest analyte concentrations with RSD <10 % [41]. The mean
259 RSD values were typically <5 % from 0.2 to 500 nM, as exemplified for Glc and GlcNAc (Fig.
260 2). For amino sugars, LOD often equaled the LOQ. However, significantly increased RSDs (20-
261 40%) were observed for Glc concentrations below 1-2 nM (Fig. 2). Thus, the effective LOQ
262 values for most identified carbohydrates ranged around 2 nM (Suppl. Table 2).

263 3.4. Analyses of lake water

264 Diverse planktonic samples from Lake Zurich were analyzed for the presence of free
265 carbohydrates and amino sugars. From a variety of tested compounds, Glc, Man, Gal, and
266 GlcNAc were readily detected in Lake Zurich. Concentrations of Glc typically ranged from 5
267 to 30 nM and the summed concentrations of Man, Gal, and Fru usually did not exceed 20
268 nM. GlcNAc was quantitatively the most important of the tested amino sugars, with a typical
269 concentration range of 0.5-2 nM. Other compounds such as Rha, Ara, Fru, ManNAc, GalNAc,
270 (GlcNAc)₂, MurNAc and some disaccharides were detected only occasionally or their
271 concentrations were below the detection limit. Even though the distribution and
272 concentrations of Glc and GlcNAc largely varied both spatially and temporally (Fig. 3), the
273 compounds were detected in the vast majority of samples. Therefore, Glc and GlcNAc were

274 selected as relevant and representative model compounds for carbohydrates and amino
275 sugars, respectively.

276 *3.5. Dynamics of Glc utilization and production in Lake Zurich*

277 To test the applicability of the adopted isotope dilution approach, Glc utilization and
278 production rates were determined in 4 incubation experiments with unfiltered water
279 samples from Lake Zurich during the summer stratification and autumn overturn periods
280 (Table 3). The concentrations of ^{13}C -labeled Glc linearly decreased over time, as illustrated in
281 Fig. 4 for 2 of the experiments, and this decline was steeper in summer samples. In contrast,
282 the concentrations of the unlabeled Glc (i.e., of the natural Glc pool) gradually increased
283 with time, which was paralleled by the increase in Glc concentrations in the controls (Fig. 4).
284 Similar temporal development of the ^{13}C -labeled and natural Glc pools were observed in the
285 other experiments (data not shown). Moreover, significant temporal changes in the ratios of
286 the ^{13}C -labeled and natural Glc resulted in pronounced changes in the total Glc pool size and
287 atom% excess ^{13}C values implying differences in Glc utilization and production rates (Fig. 4).
288 The latter also pointed at rapid recycling of Glc. While Glc utilization or production rates did
289 not correlate with depth, temperature nor with chlorophyll *a* values, both rates were
290 significantly higher during summer stratification than in the autumn overturn period (paired
291 t-test, $P < 0.01$, $n = 6$).

292 *3.6. Samples from eutrophic and dystrophic habitats*

293 To evaluate the performance of the new method on environmental samples with a
294 different matrix complexity as compared to the mesotrophic Lake Zurich, water samples
295 were collected from the eutrophic Hüttensee and the dystrophic peat bog
296 Unterrifferswilermoos (Fig. 5). The carbohydrate composition in these habitats was similar to
297 that of Lake Zurich: While Glc, Man, Gal, Fru were also detected in eutrophic Hüttensee and

298 in the peat bog, an unidentified disaccharide was present in both habitats. Concentrations of
299 amino sugars were below their detection limits during the time of sampling, apart from 5 nM
300 GlcNAc detected in the peat bog. In general, highest concentrations of the targeted analytes
301 were found in the dystrophic peat bog (up to 50 nM Glc as well as of a mix of Man, Gal, Fru)
302 followed by eutrophic and mesotrophic habitats. Most importantly, no significant
303 interferences in carbohydrate and amino sugar signal intensities as well as in background
304 noise levels were observed in any of the tested samples (Fig. 5). This suggests that the
305 detection limits of the targeted analytes by the here presented method are generally low in
306 freshwater habitats with different trophic states and matrix complexities.

307

308 **4. Discussion**

309 *4.1. Method optimization*

310 Our goal was to develop a method to simultaneously detect dissolved free
311 carbohydrates and amino sugars in freshwater samples at their expected *in situ*
312 concentrations that would be both, sufficiently sensitive and without time-consuming
313 sample preparation. We also aimed at avoiding any sample cleanup steps that could
314 compromise the quality of analyses and recovery of the analytes. Due to the high specificity
315 and the possibility of analyzing different isotopes, mass spectrometry (MS) was selected as a
316 superior detection system to the conventionally used pulse amperometric detectors [18].
317 Since our investigation focused on low-molecular weight compounds, the system of choice
318 was a triple quadrupole mass spectrometer with excellent resolution of the targeted
319 analytes. In contrast, Fourier-transform mass spectrometry, which has become increasingly
320 popular, is particularly promising for analyses of large and complex molecules [e.g., 42].

321 Application of MS was also the reason why we did not choose the well-established
322 separation of carbohydrates and amino sugars by anion-exchange chromatography under
323 basic conditions [e.g., 18,21]. The HPAEC method has been previously combined with MS
324 [e.g., 6,22,43,44] or with isotope ratio mass spectrometry [45] to determine carbohydrates
325 in a variety of samples such as bacterial cell hydrolysates, food and beverage products, plant
326 tissue or marine sediments. However, the above studies all required a desalting step or
327 additional removal of contaminants (e.g., amino acids) prior to the detection. To our
328 knowledge, only one study [6] has applied the HPAEC-MS method for the determination of
329 free carbohydrates and other low-molecular-weight compounds in freshwater samples,
330 albeit without providing sufficient details for easy reproduction: Thus, despite its apparent
331 potential, there were considerable obstacles to adopt the method by Berggren and
332 coworkers [6] for our routine analyses of free carbohydrates without extensive testing.
333 Therefore, in order to find a suitable chromatographic setup readily compatible with MS,
334 other approaches were explored, such as normal-phase chromatography (NPC) and ion-
335 exclusion chromatography (IEC).

336 While targeted analytes with different molecular weights could be well resolved by
337 NPC using a silica resin coated with an amino phase, separation of isobaric compounds was
338 not satisfactory even when modified gradient elutions with volatile buffers at different
339 concentrations were used (data not shown). In contrast, carbohydrates (Fig. 1A,B) and amino
340 sugars (Fig. 1C) could be effectively separated on a strong cation exchange resin by IEC.
341 Moreover, the applied chromatographic conditions proved to be very stable, and no
342 significant shifts in analyte retention times was observed over a 1 year period of intense
343 column use. The new protocol also allowed for higher throughput, i.e., of 3 samples per
344 hour, which is significantly faster than previously reported [9]. IEC coupled to MS permitted

345 a rapid and sensitive detection of carbohydrates and amino sugars at low nM concentrations
346 (Fig. 2). For some analytes, such as amino sugars, even sub-nM concentrations could be
347 quantified in lake water (Fig. 3B). Detection and quantification limits are, thus, comparable
348 to the threshold concentrations reported for the HPAEC-PAD method [18].

349 It should be noted that the HPAEC method may indeed allow for a better separation
350 than IEC. However, the co-eluting compounds from different classes (e.g., hexose vs deoxy-
351 hexose) could still be unambiguously discriminated by their specific transitions in our
352 approach. Thus, the incomplete chromatographic separation could be effectively
353 compensated for by MS detection. The difficulties related to the poor chromatographic
354 resolution of Man, Gal and Fru (i.e., compounds within the same class) might be
355 circumvented by using other available columns that show a baseline separation of these
356 analytes, such as the Supelcogel Pb column (Supelco).

357 4.2. *Stable isotope tracers*

358 Despite the high detection potential of our approach, it was not our intention to
359 develop a screening method for the presence of a large range of sugars in freshwaters.
360 Instead, chromatographic conditions were specifically adjusted toward a baseline separation
361 of selected compounds such as Glc and GlcNAc to be studied in more detail. In particular, we
362 aimed at combining the accurate concentration measurement of specific carbohydrates and
363 amino sugars with their *in situ* dynamics [e.g., 46,47]. We, therefore, adopted the isotope
364 dilution technique [30,33,34] to track the utilization and production rates of Glc and GlcNAc
365 in Lake Zurich at low nM concentration ranges.

366 From our incubation experiments with stable isotope tracers (Fig. 4), utilization and
367 production rates of Glc could be simultaneously determined (Table 3), thus providing
368 valuable information on the *in situ* cycling of this compound. The isotope dilution model

369 provides a reliable estimate of the cycling rates despite a negligible difference in the total
370 compound pool during the incubation period [48]. This allows for the estimation of rates
371 even at low nM concentration range. To date, knowledge about the *in situ* dynamics of
372 particular substrate in freshwaters largely originates from determining its concentration and
373 uptake rate only [e.g., 8,20,49]. However, our incubation experiments suggest that the *in*
374 *situ* pool of dissolved free Glc in Lake Zurich is not at steady state conditions. This
375 emphasizes the importance of measuring both, uptake and production rates to understand
376 the cycling of a particular compound. Furthermore, the model uses the isotope atom%
377 excess values in order to determine whether and to which extent the compound was turned
378 over. This is particularly important since the uptake of a given compound does not
379 necessarily imply its recycling. For instance, performing parallel experiments with ¹³C-labeled
380 GlcNAc revealed a rapid utilization of this compound, whereas no measurable production of
381 GlcNAc could be detected (Suppl. Fig. 1). Thus, despite comparable utilization rates, Glc and
382 GlcNAc revealed distinctly different *in situ* dynamics.

383 The concentrations of dissolved free GlcNAc likely corresponded to ~1% of the
384 expected total amino sugar pool size in Lake Zurich, as deduced from the amino sugar
385 concentrations in the total organic matter pools of two other Swiss lakes [50]. However, this
386 is not surprising since amino sugars are principal components of structural biopolymers such
387 as peptidoglycans or chitin that are typically found in the particulate fraction of organic
388 matter, for instance, in bacterial, cyanobacterial, and algal cell walls or in zooplankton
389 exoskeleton [51-54]. Since the focus of our study was exclusively on dissolved compounds
390 that would be readily available to free-living microbial cells, these organic particles were
391 removed by filtration prior to analyses. Overall, the use of stable isotope tracers together
392 with accurate measurements of concentrations by MS allows for a highly specific and

393 sensitive analysis of the *in situ* dynamics and microbial uptake of various dissolved
394 compounds and provides a convenient alternative to the radiolabeled incubation assays.

395 4.3. Pitfalls

396 Incubation studies with ^{13}C -labeled tracers at close to *in situ* (nM) concentrations are
397 delicate, and it is recommended to tightly control the sample handling, filtration and storage
398 to prevent sample contamination or degradation. The samples should not be exposed to
399 large temperature changes, which can significantly affect microbial processes such as
400 substrate utilization rates or carbon metabolism [e.g., 55,56]. The incubations should be
401 started as soon as possible (preferably within 1 h) after sampling, because the composition
402 and activity of the microbial assemblage may rapidly change after being disconnected from
403 the natural environment [e.g., 57,58]. Finally, since the utilization and production rates are
404 calculated from the change in concentrations of natural and labeled compounds over time, a
405 linear relationship between the latter parameters is required. Thus, in order to properly
406 determine atom% excess of ^{13}C , incubation time and sampling frequency should be adjusted
407 accordingly.

408 4.4. Conclusions

409 The here described new procedure is characterized by rapid sample preparation and
410 highly accurate quantification by MS. Using direct sample injections, a range of dissolved
411 free carbohydrates and amino sugars can be detected and separated at low nM
412 concentrations in a variety of freshwater samples including eutrophic and dystrophic
413 habitats (Fig. 5). Short-termed incubation experiments with stable isotope tracers provide a
414 suitable tool for studying the microbial substrate processing.

415

416 **Acknowledgments**

417 We thank E. Loher and M. Steinkellner for their help with field sampling and M. Scherer (AB
 418 Sciex, Switzerland) for the introduction into mass spectrometry. We also thank three
 419 anonymous reviewers for their constructive criticisms of the original submission. Funding for
 420 this work was provided by the Swiss National Science Foundation under the project 31003A-
 421 141166.

422

423 **References**

- 424 [1] B. Riemann, M. Søndergaard, L. Persson, L. Johansson, Carbon metabolism and community
 425 regulation in eutrophic, temperate lakes, in: B. Riemann, M. Søndergaard (Eds.), Carbon
 426 dynamics in eutrophic, temperate lakes, Elsevier, Amsterdam, 1986, p. 267-280.
- 427 [2] F. Azam, T. Fenchel, J.G. Field, J.S. Gray, L.A. Meyer-Reil, F. Thingstad, The ecological role of
 428 water-column microbes in the sea, *Mar. Ecol. Prog. Ser.* 10 (1983) 257-263.
- 429 [3] L.J. Tranvik, M.G. Höfle, Bacterial growth on mixed cultures on dissolved organic carbon from
 430 humic and clear waters, *Appl. Environ. Microbiol.* 53 (1987) 482-488.
- 431 [4] S.R. Carpenter, J.J. Cole, M.L. Pace, M. Van De Bogert, D.L. Bade, D. Bastviken, C.M. Gille, J.R.
 432 Hodgson, J.F. Kitchell, E.S. Kritzberg, Ecosystem subsidies: terrestrial support of aquatic food
 433 webs from C-13 addition to contrasting lakes, *Ecology* 86 (2005) 2737-2750.
- 434 [5] J.J. Cole, S.R. Carpenter, M.L. Pace, M.C. Van de Bogert, J.L. Kitchell, J.R. Hodgson, Differential
 435 support of lake food webs by three types of terrestrial organic carbon, *Ecology Letters* 9
 436 (2006) 558-568.
- 437 [6] M. Berggren, H. Laudon, M. Haei, L. Ström, M. Jansson, Efficient aquatic bacterial metabolism
 438 of dissolved low-molecular-weight compounds from terrestrial sources, *ISME J.* 4 (2010) 408-
 439 416.
- 440 [7] R. Benner, K. Kaiser, Abundance of amino sugars and peptidoglycan in marine particulate and
 441 dissolved organic matter, *Limnol. Oceanogr.* 48 (2003) 118-128.
- 442 [8] L.J. Tranvik, N.O.G. Jørgensen, Colloidal and dissolved organic matter in lake water:
 443 Carbohydrate and amino acid composition, and ability to support bacterial growth.,
 444 *Biogeochemistry* 30 (1995) 77-97.
- 445 [9] C. Panagiotopoulos, R. Sempéré, Analytical methods for the determination of sugars in
 446 marine samples: A historical perspective and future directions, *Limnol. Oceanogr-Meth.* 3
 447 (2005) 419-454.
- 448 [10] M. Dubois, K.A. Gilles, J.K. Hamilton, P.A. Rebers, F. Smith, Colorimetric method for
 449 determination of sugar and related substances, *Analytical Chemistry* 28 (1956) 350-356.
- 450 [11] S.V. Mykkestad, E. Skånøy, S. Hestmann, A sensitive method for analysis of dissolved mono-
 451 and polysaccharides in seawater, *Analytical Chemistry* 56 (1997) 279-286.
- 452 [12] E. Sawicki, T.R. Hauser, T.W. Stanley, W. Elbert, The methyl-2-benzothiazolone hydrazone
 453 test, *Anal. Chem.* 33 (1961) 93-96.
- 454 [13] K. Mopper, Improved chromatographic separation on anion-exchange resins. III. Sugars in
 455 borate medium, *Anal. Biochem.* 87 (1978) 162-168.
- 456 [14] K. Mopper, R. Dawson, G. Liebezeit, H.P. Hansen, Borate complex ion exchange
 457 chromatography with fluorimetric detection for determination of saccharides, *Anal. Chem.*
 458 52 (1980) 2018-2022.

- 459 [15] A. Meyer, C. Raba, K. Fischer, Ion-pair RP-HPLC determination of sugars, amino sugars and
460 uronic acids after derivatization with p-amino benzoic acid, *Anal. Chem.* 73 (2001) 2377-
461 2382.
- 462 [16] G.L. Cowie, J.I. Hedges, Determination of neutral sugars in plankton, sediments, and wood by
463 capillary gas chromatography of equilibrated isomeric mixtures, *Anal. Chem.* 56 (1984) 497-
464 504.
- 465 [17] H. Sakugawa, N. Handa, Chemical studies on dissolved carbohydrates in water samples
466 collected from North Pacific and Bering Sea., *Oceanol. Acta* 8 (1985) 185-196.
- 467 [18] K. Mopper, C.A. Schultz, L. Chevolut, C. Germain, R. Revueta, R. Dawson, Determination of
468 sugars in unconcentrated seawater and other natural waters by liquid chromatography and
469 pulsed amperometric detection, *Environ. Sci. Technol.* 26 (1992) 133-138.
- 470 [19] K. Kaiser, R. Benner, Determination of amino sugars in environmental samples with high salt
471 content by high-performance anion-exchange chromatography and pulsed amperometric
472 detection *Anal. Chem.* 72 (2000) 2566-2572.
- 473 [20] N.O.G. Jørgensen, R.E. Jensen, Microbial fluxes of free monosaccharides and total
474 carbohydrates in freshwater determined by PAD-HPLC, *FEMS Microb. Ecol.* 14 (1994) 79-94.
- 475 [21] X. Cheng, L.A. Kaplan, Simultaneous analyses of neutral carbohydrates and amino sugars in
476 freshwaters with HPLC-PAD, *J. Chromatogr. Sci.* 41 (2003) 1-5.
- 477 [22] D.S. Wunschel, K.F. Fox, A. Fox, M.L. Nagpal, K. Kim, G.C. Stewart, M. Shahgholi, Quantitative
478 analysis of neutral and acidic sugars in whole cell hydrolysates using high-performance anion-
479 exchange liquid chromatography-electrospray ionization tandem mass spectrometry, *J.*
480 *Chromatogr. A* 776 (1997) 205-219.
- 481 [23] B. Meon, F. Jüttner, Concentrations and dynamics of free mono- and oligosaccharides in a
482 shallow eutrophic lake measured by thermospray mass spectrometry, *Aquat. Microb. Ecol.*
483 16 (1999) 281-293.
- 484 [24] C. Alonso, M. Zeder, C. Piccini, D. Conde, J. Pernthaler, Ecophysiological differences of
485 betaproteobacterial populations in two hydrochemically distinct compartments of a
486 subtropical lagoon, *Environ. Microbiol.* 11 (2009) 867-876.
- 487 [25] U. Buck, H.-P. Grossart, R. Amann, J. Pernthaler, Substrate incorporation patterns of
488 bacterioplankton populations in stratified and mixed waters of a humic lake, *Environ.*
489 *Microbiol.* 11 (2009) 1854-1865.
- 490 [26] R.R. Wright, J.E. Hobbie, Use of glucose and acetate by bacteria and algae in aquatic
491 ecosystems, *Ecology* 47 (1966) 447-464.
- 492 [27] C. Bunte, M. Simon, Bacterioplankton turnover of dissolved free monosaccharides in a
493 mesotrophic lake, *Limnol. Oceanogr.* 44 (1999) 1862-1870.
- 494 [28] R.M.W. Amon, R. Benner, Rapid cycling of high-molecular-weight dissolved organic matter in
495 the ocean, *Nature* 369 (1994) 549-552.
- 496 [29] D.L. Kirchman, Y. Suzuki, C. Garside, H.W. Ducklow, High turnover rates of dissolved organic
497 carbon during a spring phytoplankton bloom, *Nature* 352 (2001) 612-614.
- 498 [30] O. Köster, F. Jüttner, NH_4^+ utilization and regeneration rates in freshwater lakes determined
499 by GC-MS of derivatized dihydroindophenol, *J. Microbiol. Meth.* 37 (1999) 65-76.
- 500 [31] T. Hama, K. Yanagi, Production and neutral aldose composition of dissolved carbohydrates
501 excreted by natural marine phytoplankton populations, *Limnol. Oceanogr.* 46 (2001) 1945-
502 1955.
- 503 [32] T. Hasegawa, H. Fukuda, I. Koike, Effects of glutamate and glucose on N cycling and the
504 marine plankton community, *Aquat. Microb. Ecol.* 41 (2005) 125-130.
- 505 [33] T.H. Blackburn, Method for measuring rates of NH_4^+ turnover in anoxic marine sediments,
506 using a $^{15}\text{N-NH}_4^+$ dilution technique, *Appl. Environ. Microb.* 37 (1979) 760-765.
- 507 [34] J. Caperon, D. Schell, J. Hirota, E. Laws, Ammonium excretion rates in Kaneohe Bay, Hawaii,
508 measured by a ^{15}N isotope dilution technique, *Mar. Biol.* 54 (1979) 33-40.

- 509 [35] P. Bossard, S. Gammeter, C. Lehmann, F. Schanz, R. Bachofen, H.-R. Bürgi, D. Steiner, U.
510 Zimmermann, Limnological description of the Lakes Zürich, Lucerne, and Cadagno, *Aquat. Sci.*
511 63 (2001) 225-249.
- 512 [36] I.D. Cuthbert, P. del Giorgio, Toward a standard method of measuring color in freshwater,
513 *Limnol. Oceanogr.* 37 (1992) 1319-1326.
- 514 [37] J.C. Tran, A.A. Doucette, Cyclic polyamide oligomers extracted from nylon 66 membrane filter
515 disks as a source of contamination in liquid chromatography/mass spectrometry, *J. Am. Soc.*
516 *Mass Spectr.* 17 (2006) 652-656.
- 517 [38] M. Lindenberg, C. Wiegand, J.B. Dressman, Comparison of the adsorption of several drugs to
518 typical filter materials, *Dissolut. Technol.* 12 (2005) 22-25.
- 519 [39] F. Jüttner, B. Meon, O. Köster, Quasi *in situ* separation of particulate matter from lakewater
520 by hollow-fibre filters to overcome errors caused by short turnover times of dissolved
521 compounds, *Water Res.* 31 (1997) 1637-1642.
- 522 [40] A. Shrivastava, V.B. Gupta, Methods for the determination of limit of detection and limit of
523 quantification of the analytical methods, *Chron. Young Sci.* 2 (2011) 21-25.
- 524 [41] D.A. Armbruster, T. Pry, Limit of blank, limit of detection and limit of quantification, *Clin.*
525 *Biochem. Rev.* 29(Suppl 1) (2008) S49-S52.
- 526 [42] K.D. Henry, E.R. Williams, B.H. Wang, F.W. McLafferty, J. Shabanowitz, D.F. Hunt, Fourier-
527 transform mass spectrometry of large molecules by electrospray ionization, *P. Natl. Acad. Sci.*
528 *USA* 86 (1989) 9075-9078.
- 529 [43] C. Bruggink, R. Maurer, H. Herrmann, S. Cavalli, F. Hoefler, Analysis of carbohydrates by anion
530 exchange chromatography and mass spectrometry, *J. Chromatogr. A* 1085 (2005) 104-109.
- 531 [44] T. Kirchgorg, S. Schüpbach, N. Kehrwald, D.B. McWethy, C. Barbante, Method for the
532 determination of specific molecular markers of biomass burning in lake sediments, *Org.*
533 *Geochem.* 71 (2014) 1-6.
- 534 [45] H.T.S. Boschker, T.C.W. Moerdijk-Poortvliet, P. van Breugel, M. Houtekamer, J.J. Middelburg,
535 A versatile method for stable carbon isotope analysis of carbohydrates by high-performance
536 liquid chromatography/isotope ratio mass spectrometry, *Rapid Commun. Mass Sp.* 22 (2008)
537 3902-3908.
- 538 [46] J. Nedoma, J. Vrba, J. Hejzlar, K. Šimek, V. Straškrabová, N-acetylglucosamine dynamics in
539 freshwater environments: Concentration of amino sugars, extracellular enzyme activities,
540 and microbial uptake, *Limnol. Oceanogr.* 39 (1994) 1088-1100.
- 541 [47] H.-P. Grossart, A. Engel, C. Arnosti, C.L. De La Rocha, A.E. Murray, U. Passow, Microbial
542 dynamics in autotrophic and heterotrophic seawater mesocosms. III. Organic matter fluxes,
543 *Aquat. Microb. Ecol.* 49 (2007) 143-156.
- 544 [48] E. Laws, Isotope dilution models and the mystery of vanishing ¹⁵N, *Limnol. Oceanogr.* 29
545 (1984) 379-386.
- 546 [49] M. Weiss, M. Simon, Consumption of labile dissolved organic matter by limnetic
547 bacterioplankton: the relative significance of amino acids and carbohydrates, *Aquat. Microb.*
548 *Ecol.* 17 (1999) 1-12.
- 549 [50] D. Carstens, K.E. Köllner, H. Bürgmann, B. Wehrli, C.J. Schubert, Contribution of bacterial cells
550 to lacustrine organic matter based on amino sugars and D-amino acids, *Geochim.*
551 *Cosmochim. Acta* 89 (2012) 159-172.
- 552 [51] J. Blackwell, K.D. Parker, K.M. Rudall, Chitin fibers of the diatom *Thalassiosira fluviatilis* and
553 *Cyclotella cryptica*, *J. Molec. Biol.* 28 (1967) 383-385.
- 554 [52] I.W. Sutherland, Biosynthesis and composition of gram-negative bacterial extracellular and
555 wall polysaccharides, *Annu. Rev. microbiol.* 39 (1985) 243-270.
- 556 [53] E. Cohen, Chitin biochemistry: synthesis and inhibition, *Annu. Rev. Entomol.* 32 (1987) 71-93.
- 557 [54] C. Bertocchi, L. Navarini, A. Cesàro, Polysaccharides from cyanobacteria, *Carbohydr. Polym.*
558 12 (1990) 127-153.

- 559 [55] J.K. Apple, P. Del Giorgio, W.M. Kemp, Temperature regulation of bacterial production,
 560 respiration, and growth efficiency in a temperate salt-marsh estuary, *Aquat. Microb. Ecol.* 43
 561 (2006) 243-254.
- 562 [56] C.C. Crawford, J.E. Hobbie, K.L. Webb, The utilization of dissolved free amino acids by
 563 estuarine microorganisms, *Ecology* 55 (1974) 551-563.
- 564 [57] E.B. Sherr, B.F. Sherr, C.T. Sigmon, Activity of marine bacteria under incubated and *in situ*
 565 conditions, *Aquat. Microb. Ecol.* 20 (1999) 213-223.
- 566 [58] R. Massana, C. Pedrós-Alió, E.O. Casamayor, J.M. Gasol, Changes in marine bacterioplankton
 567 phylogenetic composition during incubations designed to measure biogeochemically
 568 significant parameters, *Limnol. Oceanogr.* 46 (2001) 1181-1188.

569 **Figures captions**

570 Fig. 1. LC-MS extracted ion chromatograms of a mixture (100 nM each) of hexoses and
 571 deoxy-hexoses (A), disaccharides and pentoses (B), and amino sugars (C) acquired using a
 572 multiple reaction monitoring (MRM) scan mode. Identical transitions were applied for
 573 isobaric analytes. Glc - glucose, Man - mannose, Gal - galactose, Fru - fructose, Fuc - fucose,
 574 Rha - rhamnose, Scl – sucralose, Cel - cellobiose, Mal - maltose, Suc - sucrose, Xyl - xylose,
 575 Lyx - lyxose, Ara - arabinose, Rib - ribose, MurNAc - N-acetylmuramic acid, (GlcNAc)₂ - N,N'-
 576 diacetylglucosamine, ManNAc - N-acetyl-D-mannosamine, GlcNAc - N-acetyl-D-glucosamine,
 577 GalNAc - N-acetyl-D-galactosamine.

578

579 Fig. 2. Calibration curves of Glc (A) and GlcNAc (B) to determine the limits of detection (LOD)
 580 and limits of quantification (LOQ) in samples from Lake Zurich. Values are means of 6

581 repetitive scans. Error bars show relative standard deviation. Arrows indicate the
582 corresponding LOD and LOQ values.

583

584 Fig. 3. Concentrations of dissolved free Glc and GlcNAc in selected samples from Lake Zurich
585 on 18 April (A) and 08 May (B) 2013. Values are means of triplicates. Error bars show
586 standard deviations. Note different scales of Glc concentrations.

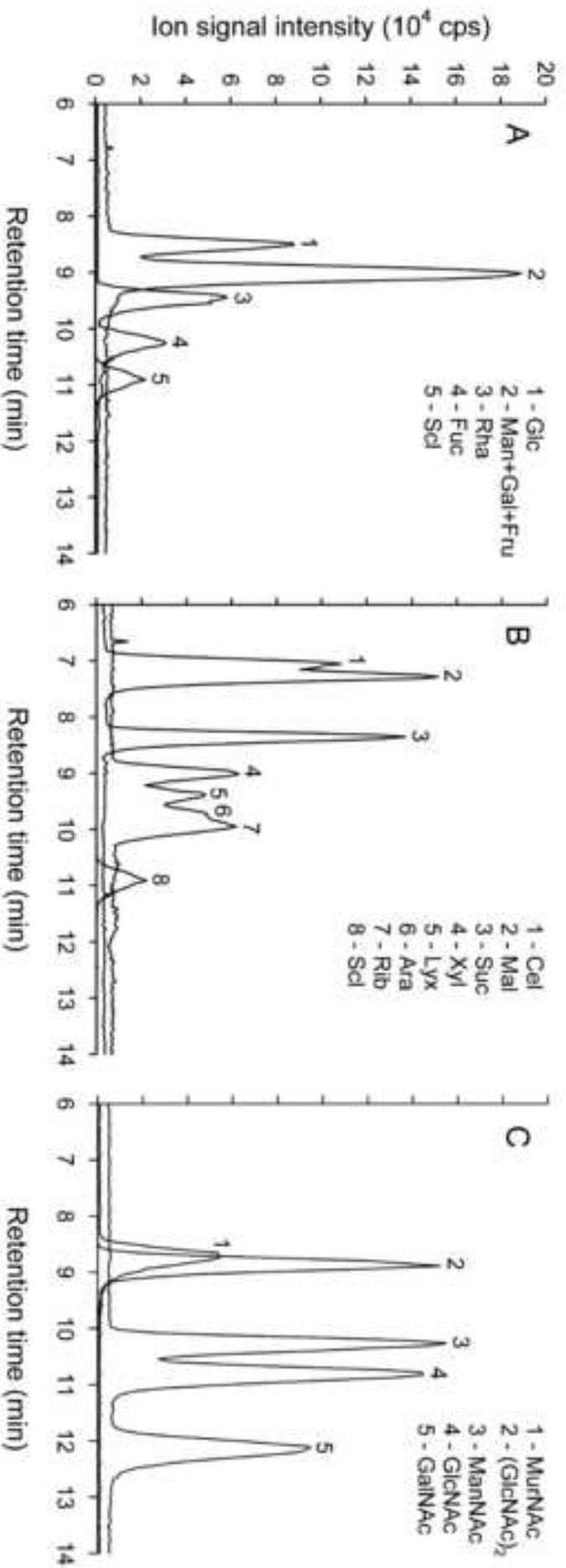
587 Fig. 4. Temporal changes in the concentrations of total, natural (^{12}C) and ^{13}C -labeled Glc in
588 selected experimental incubations performed with epilimnetic samples (6 and 12 m) from
589 Lake Zurich taken during summer stratification (A) and autumn overturn (B) periods in 2013.
590 A linear fit of the total Glc concentrations (as a sum of natural and ^{13}C -labeled Glc pools) is
591 shown. Bars represent the concentrations of Glc in the control incubations that were not
592 amended with the ^{13}C -labeled Glc. Values are means of triplicates. Error bars show standard
593 deviations (range is shown for controls). The corresponding plots of $\ln(\text{atom}\% \text{ excess } ^{13}\text{C})$
594 against $\ln(\text{Glc}_t / \text{Glc}_{t0})$ from the same incubations during summer stratification (C) and
595 autumn overturn (D). Glc_t – concentration of total Glc at different time intervals, Glc_{t0} –
596 initial total Glc concentration.
597

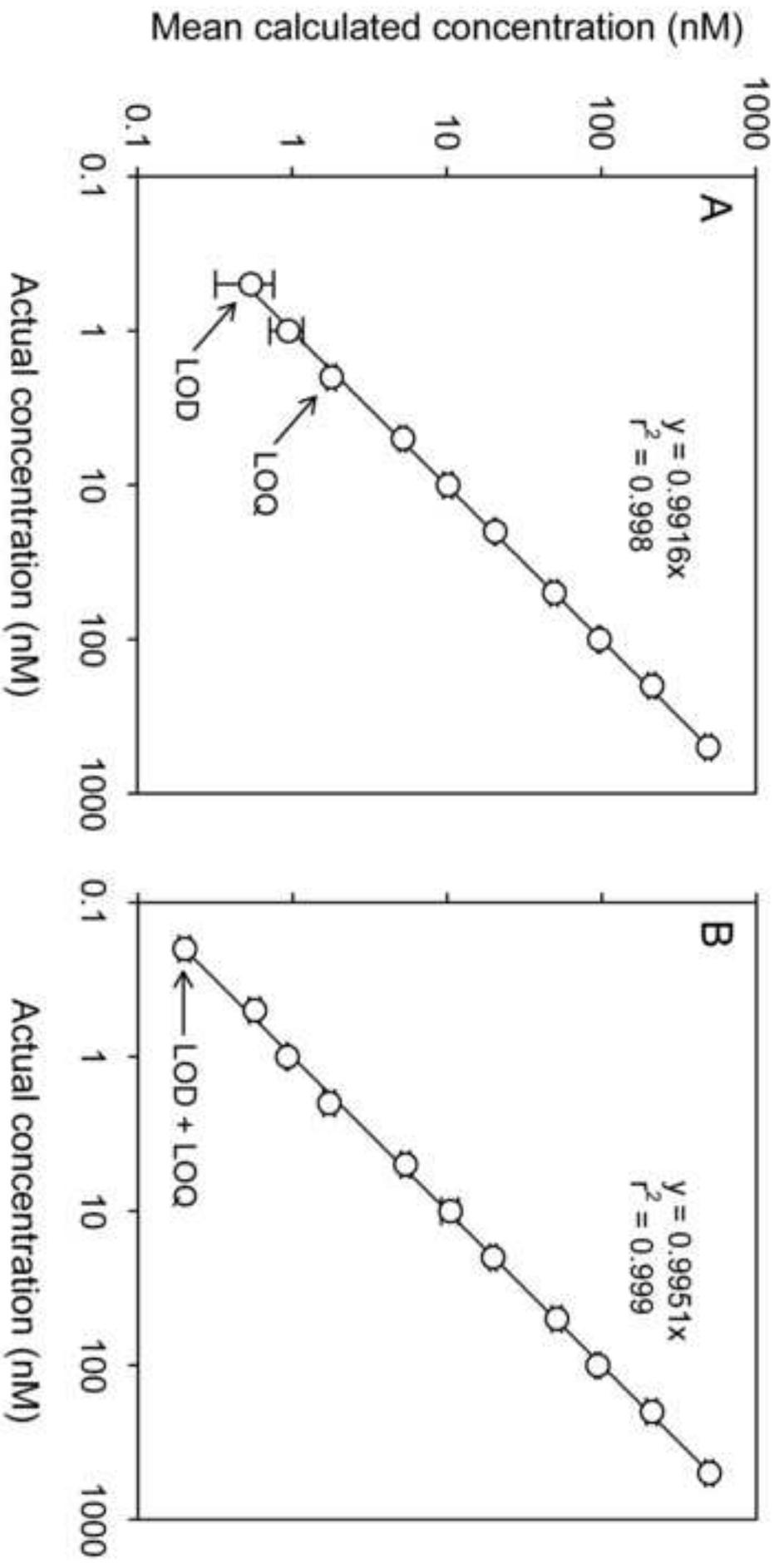
598 Fig. 5. LC-MS total ion chromatograms (TIC) of selected carbohydrates and amino sugars
599 detected in the mesotrophic Lake Zurich (A), eutrophic Hüttensee (B), and dystrophic
600 Unterrifferswilermoos peat bog (C) in March 2014. Peak identification: (1) unknown
601 disaccharide, (2) glucose, (3) galactose, mannose, fructose, (4) N-acetyl-glucosamine, (5)
602 sucralose (internal standard). Different mass spectrometric settings were applied to detect
603 carbohydrates, N-acetyl-amino sugars and internal standard between 0-10, 10-12, and 12-15
604 min, respectively. Note different scales of ion signal intensities.
605

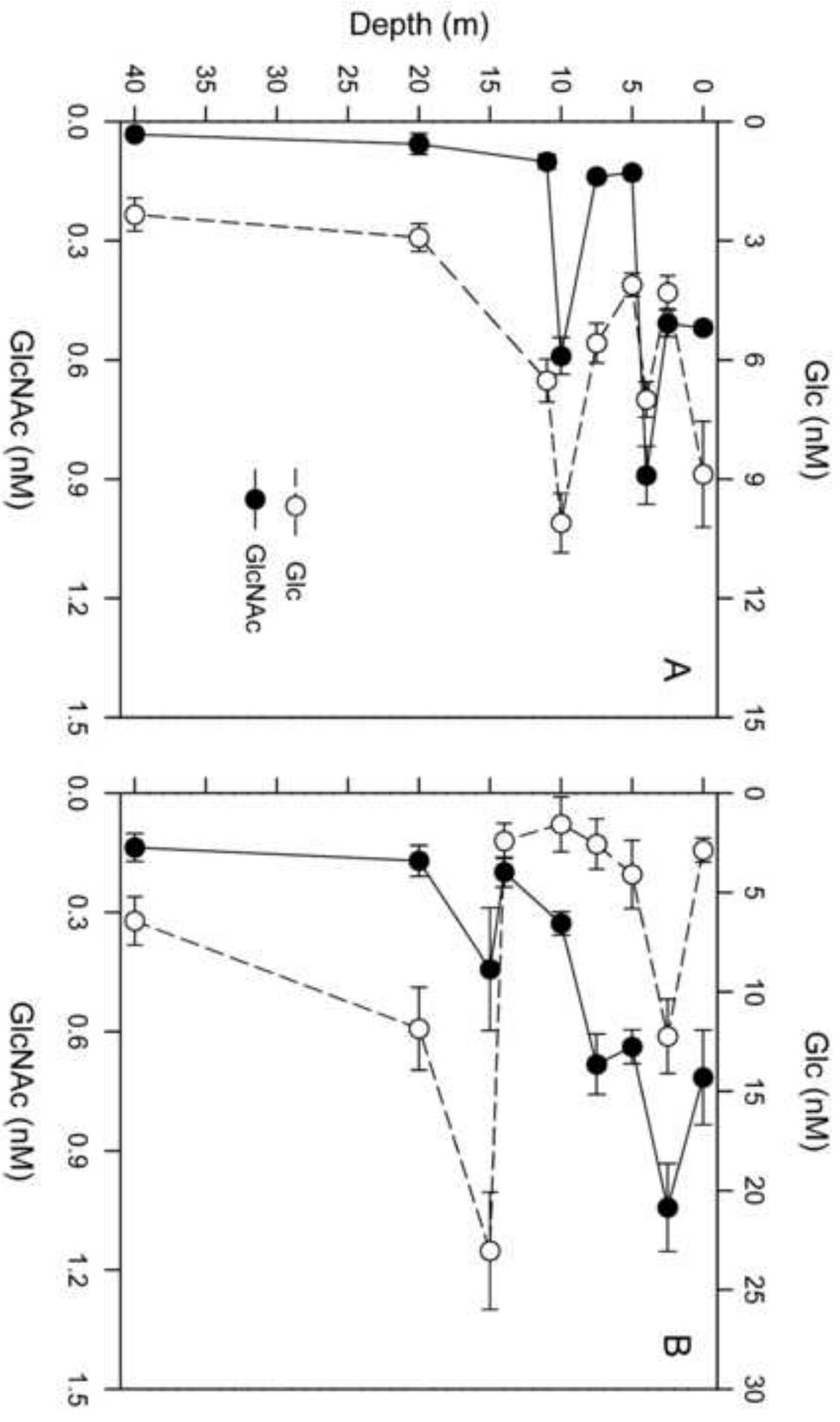
Highlights:

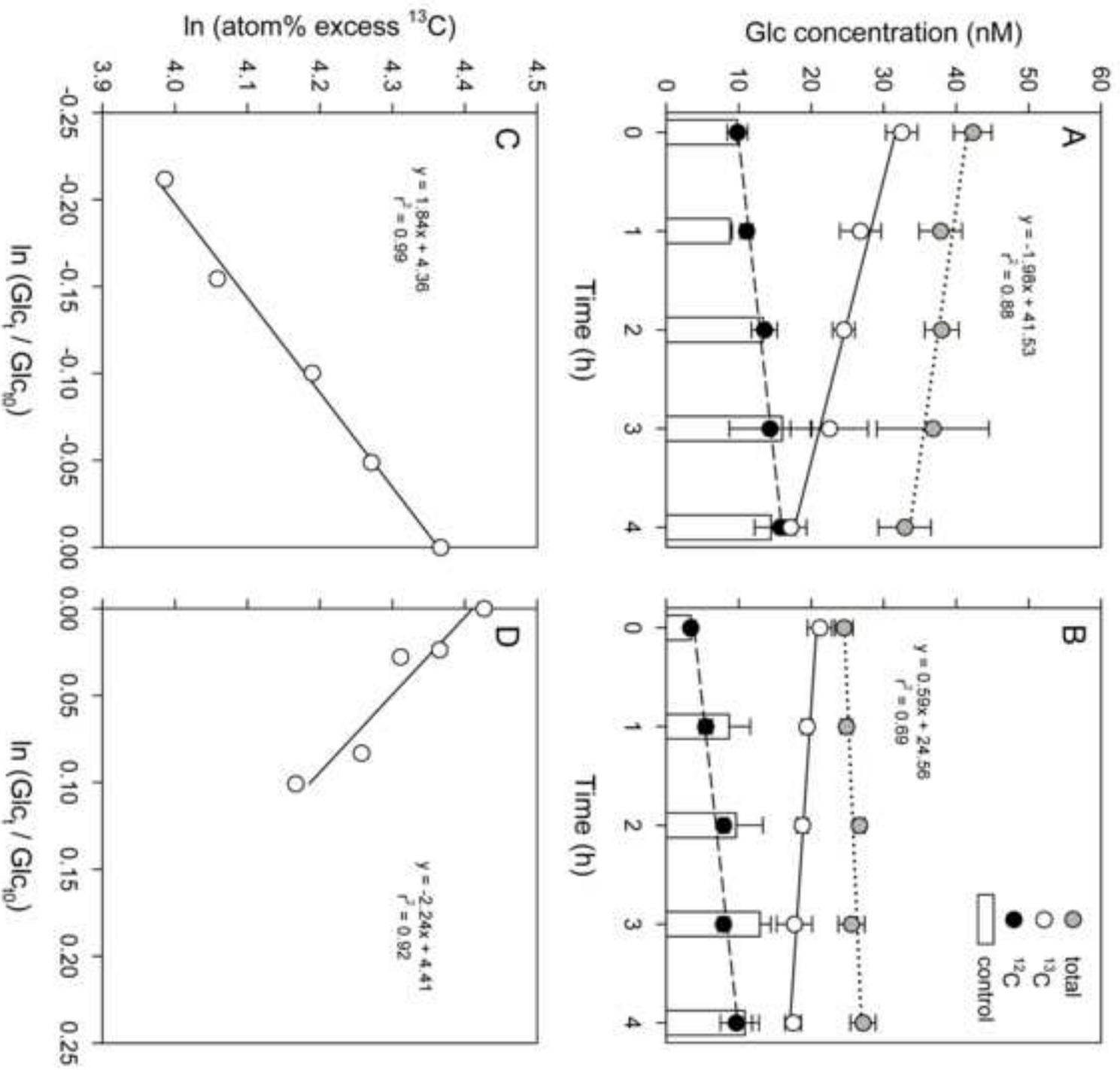
- Novel HPLC-MS method for carbohydrate and amino sugar analysis in freshwaters
- Rapid and efficient determination without derivatization or sample pre-concentration
- Implementation of stable isotope tracers to study compound biogeochemical cycling
- Glucose and N-acetyl-glucosamine showed different in situ dynamics in Lake Zurich

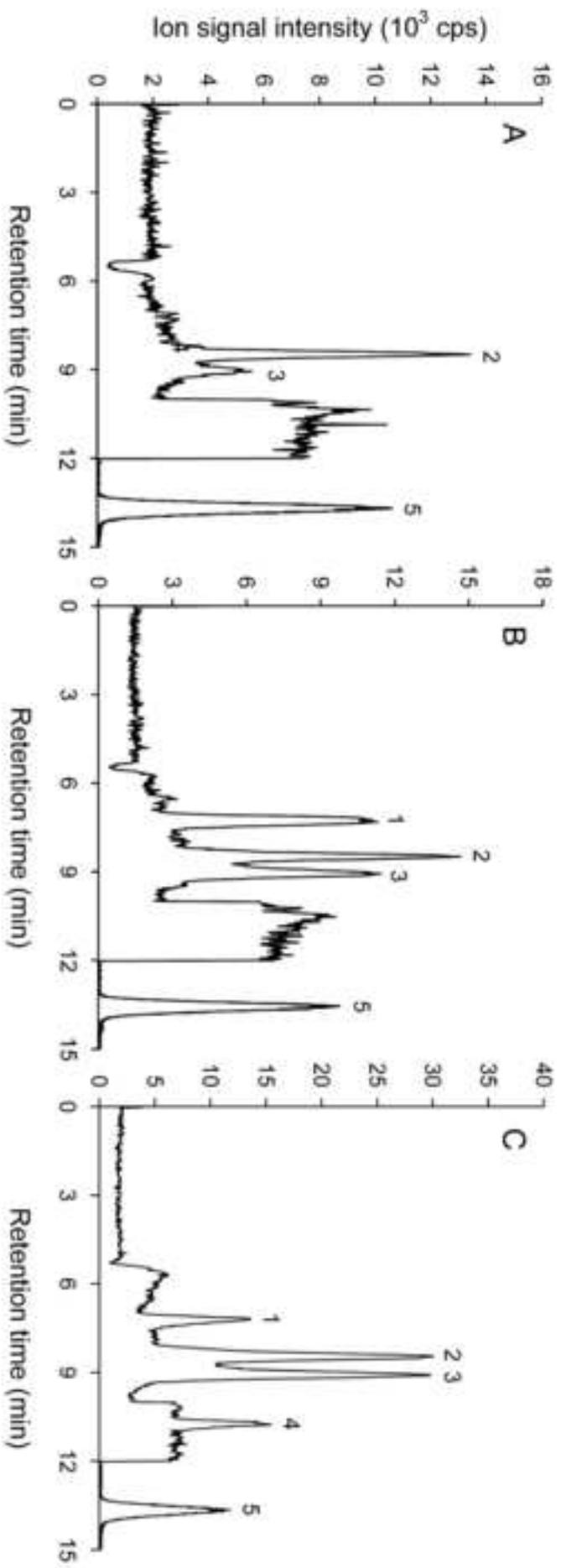
Accepted Manuscript











Tables

Table 1: Evaluation of syringe filters with different membrane types used for sample filtration prior to HPLC-MS analyses of carbohydrates and amino sugars. CA – cellulose acetate, GF – glass fiber, NY – nylon, PES – polyethersulphone, PVDF – polyvinylidene fluoride, PTFE – polytetrafluoroethylene. Recommended filters are highlighted in bold.

Membrane type	Pore size & diameter	Manufacturer
CA	0.22 μm , 25 mm	Whatman
GF	0.22 μm , 25 mm	Tisch Scientific
NY	0.22 μm , 25 mm	Millipore
PES	0.22 μm, 13 mm	Pall
PES (hollow fiber)	0.05 & 0.22 μm	Spectrum
PVDF	0.1 μm , 25 mm	Millipore
PTFE	0.22 μm , 25 mm	Whatman

Table 2: Multiple reaction monitoring (MRM) parameters applied to detect carbohydrates and internal standard (Scl, negative ion mode) and amino sugars (positive ion mode). Transitions (pairs of precursor and product ions) and the corresponding declustering potential (DP), collision energy (CE) and cell exit potential (CXP) parameters are shown. Gal - galactose, Glc - glucose, Fru - fructose, Man - mannose, Fuc - fucose, Rha - rhamnose, Ara - arabinose, Lyx - lyxose, Rib - ribose, Xyl - xylose, Cel - cellobiose, Mal - maltose, Suc - sucrose, Scl – sucralose, GalNAc - N-acetyl-D-galactosamine, GlcNAc - N-acetyl-D-glucosamine, ManNAc - N-acetyl-D-mannosamine, (GlcNAc)₂ - N,N'-diacetylglucosamine, MurNAc - N-acetylmuramic acid.

Compound	DP (V)	Transitions (m/z)	CE (V)	CXP (V)
Gal, Glc, Fru, Man	-75	179/59	-22	-5
		179/71	-22	-9
		179/89	-14	-13
		179/119	-8	-19
Fuc, Rha	-55	163/59	-22	-5
		163/101	-12	-15
Ara, Lyx, Rib, Xyl	-75	149/59	-18	-7
		149/71	-12	-13
		149/89	-10	-7
Cel, Mal, Suc	-75	341/59	-44	-9
		341/101	-22	-9
		341/161	-8	-11
		341/179	-8	-21
Scl (internal standard)	-150	397/361	-14	-15
		397/359	-16	-19
		395/359	-10	-19
GalNAc, GlcNAc, ManNAc	36	222/126	17	14
		222/138	21	20
		222/144	21	10
		222/186	15	12
(GlcNAc) ₂	126	425/84	59	18
		425/126	35	18
		425/204	15	22
MurNAc	66	294/84	29	16
		294/114	41	14

Table 3: Dynamics of Glc in selected samples from Lake Zurich during summer stratification (August) and autumn overturn (October) periods in 2013. The corresponding *in situ* temperature, chlorophyll *a* and Glc concentrations are shown. Glc utilization and production rates were obtained from the incubation experiments with ^{13}C -labeled Glc.

Sampling date	Depth (m)	Temperature (°C)	Chl <i>a</i> ($\mu\text{g l}^{-1}$)	Glc (nM)	Utilization rate (nM Glc h $^{-1}$)	Production rate (nM Glc h $^{-1}$)
05 Aug 2013	6	25	5	9.8	3.65	1.66
	12	14	34	8.6	5.65	3.09
	20	7	4	3.5	3.33	2.94
12 Aug 2013	6	24	6	5.5	5.89	6.22
	12	13	33	9	5.14	2.76
	22	7	3	9.4	1.75	2.63
21 Oct 2013	5	15	17	3.3	0.87	1.57
	10	15	17	5.7	0.53	0.91
	40	5	0.4	3.9	1.68	1.65
28 Oct 2013	5	14	19	2.8	2.68	1.51
	10	14	17	1.5	1.96	1.24
	40	5	0.6	7	1.74	0.61